

# Bioaccumulations of heavy metals in submerged macrophytes in the mountain river Biała Łądecka (Poland, Sudety Mts.)

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**Abstract:** The study was conducted on the Biała Łądecka River which is a mountain river. It is similar to many European mountain rivers in terms of hydromorphology and catchment management. The aim of this study was to determine the bioconcentration factors of heavy metals (Pb, Cd, Hg, Ni, Cr, Cu and Zn) in *Ranunculus aquatilis* (L.) Dumort., *Fontinalis antipyretica* (L. ex Hedw.), and *Lemanea fluviatilis* (L.) C.Ag. The content of metals in water, sediment, and submerged plants was determined. The metal concentrations in plants can be arranged as follows: Hg < Cd < Cr < Ni < Cu < Pb < Zn. The highest concentrations of Hg, Ni, Cr, and Cu were observed in *F. antipyretica*, but the highest concentrations of Pb, Cd, and Zn were in *R. aquatilis*. *L. fluviatilis* always contained the least amounts of heavy metals. Bioconcentration factors (BCFs) were lowest in *L. fluviatilis* and highest in *F. antipyretica*. Among the analyzed metals, plants accumulated the highest amount of Zn, and the least of Hg. The BCFs for Zn were from 24111 (in *L. fluviatilis*) to 97574 (in *R. aquatilis*), and BCFs for Hg were from 29 (in *L. fluviatilis*) to 226 (in *F. antipyretica*).

## Introduction

Water pollution is a fact. We are looking for better (cheaper, easier, faster, more accurate) methods for assessing contamination. A particular challenge is to assess the pollution in the fast-flowing waters of mountain rivers. High flow rates result in a change in the content of pollutants in the water.

Macrophytes have proved to be good material for the study since they are easy to obtain and prepare for testing. Among the aquatic plants, submerged plants deserve special attention. Due to the physiology of submerged plants, they may accumulate the largest amounts of pollutants (Senze et al. 2009).

We wanted to make our study as universal as possible. For this reason, we chose for the study the Biała Łądecka river which is a typical European mountain river. The Biała Łądecka River is a right tributary of the Nysa Kłodzka River, with a length of 53 km and a basin area of 314.6 km<sup>2</sup>. The Biała Łądecka River begins in the Bialskie Mountains at an altitude of 1090 meters above sea level and turns into the Nysa Kłodzka River at its 133.1 km in the village of Krosnowice at 236 meters above sea level. The river basin is dominated by forests, extensive agriculture as well as urban development. Pollution is prominent because of the crystal glass factory,

municipal sewage treatment plants and numerous trout farms. The middle-part of the river (16.3 km) is designated as a Special Protected Area “PLH020035 Biała Łądecka” under the Natura 2000 network.

The study involved macrophytes commonly found in the Biała Łądecka River. There are dominant submerged species of macrophytes in this river, representing different systematic groups of plants.

*Ranunculus aquatilis* (L.) Dumort. is a species representing a group of *Angiosperms* Cronquist., of the family *Ranunculaceae* (Juss.). *Ranunculus aquatilis* (L.) Dumort. is a perennial plant and hydrophyte. This plant blooms from April to August. It occurs in shallow and slow-flowing water and canals (Cook 1963, Dahlgren 1991). *Ranunculus* (L.) subgenus *Batrachium* (DC.) A. Gray – *Batrachium* was first treated as a subgenus by A. Gray in 1886 but the rank has varied from section to genus (Dahlgren, 1991). Cook (1963) found the subgenus rank to be the most appropriate. The subgenus is mainly located in Atlantic European countries. All species except two occur in Europe. The subgenus *Batrachium* has 17 species (Dahlgren 1991).

*Fontinalis antipyretica* (L. ex Hedw.) is a taxon representing the subclass *Bryidae*. It occurs in Europe, North Asia, northern Africa and North America, in clear-flowing or

standing waters. This moss is common throughout Poland. It grows on underwater rocks, stones, and pieces of wood in shady places. *F. antipyretica* is often found in temperate regions, in flowing freshwater streams and ponds. Of the 20 species of water moss, 18 are native to North America. A brook moss may have shoots 30 to 100 (rarely up to 200) cm long and is usually attached to a stone or a tree root. The most common species is *F. antipyretica*. This species has long, slender branches covered with glossy, yellowish green or dark green phyllids (leaves), 4 to 7 mm long and arranged in three ranks. Male and female reproductive organs are on separate plants (Atherton et al. 2010).

*Lemanea fluviatilis* (L.) C.Ag. is a species representing the division *Rhodophyta*, of the order *Batrachospermales* (Bory de Saint-Vincent 1808). This taxon occurs in Europe (Kučera et al. 2004), in North America (Vis et al. 1992), South America and Australia. The freshwater red algae of the family *Lemaneaceae* are characterized by an uniaxial cartilaginous and pseudoparenchymatous gametophyte thallus with internal carposporophytes (Vis et al. 1992). In Poland, the species occurs in several places especially in the southern part of the country (Starmach 1977). There is only a small-number of locations with the species *L. fluviatilis* since it needs specific ecological conditions. For this reason, this alga is on the red list, in those countries where this species is defined as vulnerable (Siemińska 2006).

The first aim of this study was to determine the bioconcentration factors (BCFs) for heavy metals (Pb, Cd, Hg, Ni, Cr, Cu and Zn) in *Ranunculus aquatilis* (L.) Dumort., *Fontinalis antipyretica* (L. ex Hedw.), and *Lemanea fluviatilis* (L.) C.Ag. The second aim was to investigate whether the bioconcentration factors depend on basic physicochemical

parameters of water as well as the concentration of heavy metals in water and sediment.

### Materials and methods

The research was conducted in 2005, in the section of the river that runs from the mouth (in Krosnowice) to the village of Stary Gieraltów. Samples of water, sediment, and plants were taken at 9 points (Fig 1), three times a year (in the spring, summer and autumn). Coordinates of sampling sites were determined using GPS map 76CS (GARMIN).

The following physicochemical parameters were determined: temperature (°C), electrolytic conductivity ( $\mu\text{S} \cdot \text{cm}^{-1}$ ), water pH, dissolved oxygen ( $\text{mgO}_2 \cdot \text{dm}^{-3}$ ), water alkalinity ( $\text{mgCaCO}_3 \cdot \text{dm}^{-3}$ ), water hardness ( $\text{mgCaCO}_3 \cdot \text{dm}^{-3}$ ), and the biochemical oxygen demand over 5 days –  $\text{BOD}_5$  ( $\text{mgO}_2 \cdot \text{dm}^{-3}$ ).

The sediment samples and plants were “wet” mineralized in a pressure microwave oven MARS 5 (CEM Corporation, USA). The sediment samples were mineralized with a mixture of  $\text{HNO}_3$  and  $\text{HClO}_4$  at 3÷1, and the plants with  $\text{HNO}_3$ .

The amount of metals in mineralizate were determined using atomic absorption spectrophotometry (AAS). The flame method (SpectrAA 220 – Varian, Australia) was used to determine Pb, Cd, Ni, Cr, Cu, and Zn. The cold vapor method (TMA 254 – Tesla, the Czech Republic) was used to determine Hg.

The results of the analyses were verified with the certified reference materials: BCR 60 (Commission of the European Communities, Community Bureau of Reference) – aquatic plant *Lagarosiphon major* and CRM 482 (Commission of the European Communities, Community Bureau of Reference)

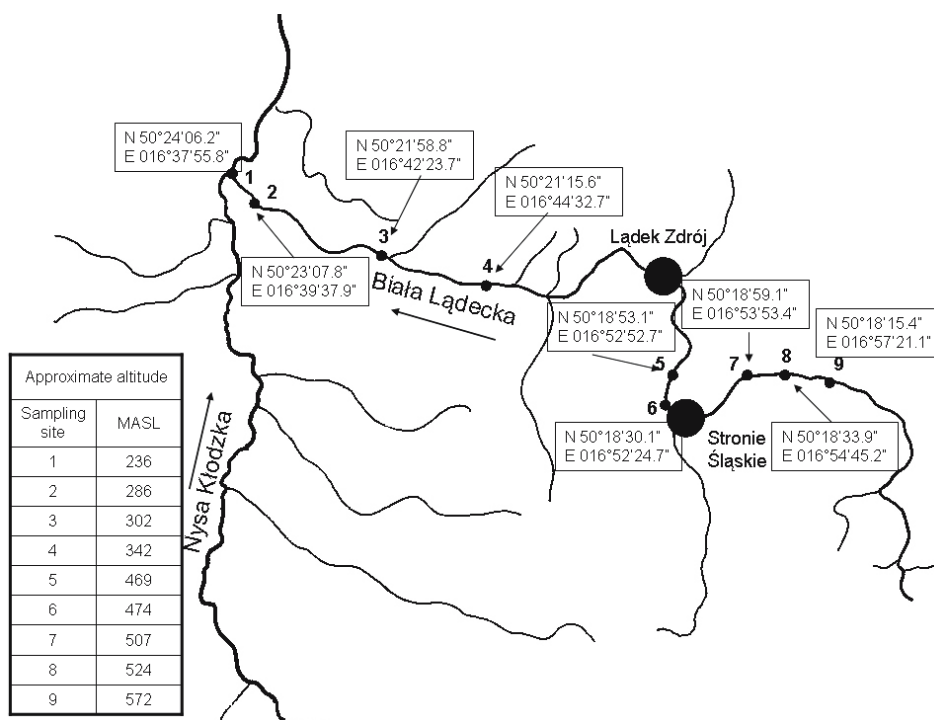


Fig. 1. Location of sampling site on Biała Łądecka River

MASL – meters above sea level

– *Pseudevernia furfuracea*, LGC 6187 (Laboratory of the Government Chemistry) – river sediments. Certified values for the reference materials used and the obtained concentrations (5 replicates) are summarized in Table 1.

The accuracy of the method was understood as the difference between the real value of the studied indicator, and the average value of the obtained measurement that was below 5%.

Metal concentrations in the deposits and metal concentrations in the plants are presented in  $\text{mg} \cdot \text{kg}^{-1}$  of dry weight, in water in  $\mu\text{g} \cdot \text{dm}^{-3}$ .

The bioconcentration factors (BCFs) were calculated as the relationship of metal concentration in the plant to the concentration of metal in water.

The statistical analysis of the results (the minima, the maxima, the mean, the standard deviation, differences, correlations) was carried out using Statistica 10.0 (StatSoft).

## Results

The recorded physical and chemical indicators of water quality are summarized in Table 2. According to the current classification in Poland, the physical and chemical water

parameters in the tested section, classified the Biała Łądecka River as having a class I freshwater quality.

The pH of the water of the Biała Łądecka River was slightly alkaline (pH 6.9–8.1). Electrolytic conductivity of the water and its total hardness were relatively low. The content of dissolved  $\text{O}_2$  was high, which is typical for the fast-running and stirred-up water of mountain streams and rivers. Average  $\text{BOD}_5$  was low, but its maximum value was high. These results may indicate water contamination with organic substances.

The metal contents which were found in the water of the Biała Łądecka River are shown in Table 3. Different levels of metals were found in the water from  $0.57 \mu\text{g} \cdot \text{dm}^{-3}$  (Hg) to  $8.43 \mu\text{g} \cdot \text{dm}^{-3}$  (Zn). According to Kabata-Pendias et al. (1993), Pb, Ni, Cr, Cu, and Zn concentrations are characteristic for Polish rivers, while concentrations of Hg and Zn indicate contamination.

The contents of five metals – Pb, Hg, Ni, Cu i Zn were examined in sediment (Table 4). The concentration of Hg in the sediment was lowest ( $0.03 \text{ mg} \cdot \text{kg}^{-1}$ ). Concentrations of Cu and Ni were considered to be at the middle level ( $13.78$  and  $14.71 \text{ mg} \cdot \text{kg}^{-1}$ , respectively), while the concentration of Zn in the sediment was highest ( $74.24 \text{ mg} \cdot \text{kg}^{-1}$ ).

**Table 1.** Certified and obtained concentration of Pb, Cd, Hg, Ni, Cr, Cu and Zn in used reference materials ( $\text{mg} \cdot \text{kg}^{-1}$ )

	Pb	Cd	Hg	Ni	Cr	Cu	Zn
CRM 482 certified	40.900	0.560	0.480	2.470	4.120	7.030	100.600
CRM 482 obtained	42.082	0.576	0.494	2.485	4.222	7.247	102.219
SD	0.902	0.065	0.340	0.148	0.190	0.525	2.488
difference (%)	2.89	2.78	2.92	0.61	2.48	3.08	1.61
BCR 60 certified	63.800	2.120	0.340	40.000*	26.000*	51.200	313.000
BCR 60 obtained	64.027	2.160	0.338	40.768	26.850	51.650	311.322
SD	2.308	0.179	0.030	1.942	2.399	2.670	3.721
difference (%)	0.36	1.89	-0.53	1.92	3.27	0.87	-0.54
LGC 6187 certified	77.200	2.700	1.400	34.700	84.000	–	439.000
LGC 6187 obtained	78.100	2.746	1.424	34.740	84.836	–	438.690
SD	1.122	0.128	0.037	1.514	1.791	–	1.758
difference (%)	1.16	1.70	1.74	0.12	1.00	–	-0.07

\* – not certificated values; SD – standard deviation

**Table 2.** Physical and chemical parameters of water – temperature ( $^{\circ}\text{C}$ ), pH, electrolytic conductivity ( $\mu\text{S} \cdot \text{cm}^{-1}$ ), alkalinity ( $\text{mg CaCO}_3 \cdot \text{dm}^{-3}$ ), total hardness ( $\text{mg CaCO}_3 \cdot \text{dm}^{-3}$ ), dissolved oxygen ( $\text{mg O}_2 \cdot \text{dm}^{-3}$ ) and  $\text{BOD}_5$  ( $\text{mg O}_2 \cdot \text{dm}^{-3}$ )

	Temperature	pH	Electrolytic conductivity	Alkalinity	Total hardness	Dissolved oxygen	$\text{BOD}_5$
Average	10.12	–	144.35	44.646	74.427	11.51	2.8
Min	6.5	6.9	87	19.99	55.34	10.2	1.1
Max	14.0	8.1	236	99.94	110.67	12.4	7.7
SD	2.43	–	42.83	19.05	16.93	0.60	1.22

SD – standard deviation

**Table 3.** Concentrations of Pb, Hg, Cd, Ni, Cr, Cu and Zn in water ( $\mu\text{g} \cdot \text{dm}^{-3}$ )

	Pb	Hg	Cd	Ni	Cr	Cu	Zn
Average	4.99	0.57	0.79	1.55	0.90	2.09	8.43
Min	0.60	0.20	0.20	0.10	0.10	1.00	0.10
Max	11.10	1.40	1.90	5.80	2.10	4.20	29.60
SD	2.86	0.31	0.42	1.36	0.55	0.86	6.32

SD – standard deviation

**Table 4.** Concentrations of Pb, Hg, Ni, Cu and Zn in bottom deposits ( $\text{mg} \cdot \text{kg}^{-1}$ )

	Pb	Hg	Ni	Cu	Zn
Average	56.28	0.03	14.71	13.78	74.24
Min	10.39	0.01	8.94	6.66	45.35
Max	134.41	0.08	18.94	20.37	116.58
SD	40.96	0.03	3.32	4.42	21.64

SD – standard deviation

Among the studied metals in plants, Zn was found most frequently, whereas Hg the least (Table 5), and the differences between concentration of these metals were very high. The concentration of Zn in *F. antipyretica* was 1,500 times higher than the concentration of Hg in *F. antipyretica*. In *R. aquatilis* the concentration of Zn was 10,300 times higher than the concentration of Hg. The significance of difference between concentrations of metals in different plant species was also calculated (Table 5).

The noted metal content in water and plants allowed for the calculation of the bioconcentration factors (BCFs). The obtained results are shown in Fig 2 and in Table 6. The trends, which are visible on the graph (Fig. 2), have been confirmed by the demonstration of the statistical significance of difference as shown in Table 6 (all metals except Pb).

The correlation between physicochemical parameters of water, metal concentrations in water, sediment, plants and the metal concentration (Tables 7, 8, and 9), and BCFs for metals in plants (Tables 10, 11, and 12) were calculated. Only statistically significant ( $p \leq 0.05$ ) correlations are presented in the Tables.

All other correlations should be considered as strong,  $r$  ranged from -1.0 to -0.5, and from 0.5 to 1.0.

## Discussion

Kozubek et al. (2002) conducted research on the Bystrzyca Dusznicka River which is similar to and located near the Biała Łądecka.

These authors found that pH and conductivity of the Bystrzyca Dusznicka River were similar to those of the Biała Łądecka. They also found that  $\text{BOD}_5$  was significantly lower ( $0.82 \text{ mg O}_2 \cdot \text{dm}^{-3}$ ), dissolved oxygen levels ( $9.68 \text{ mg O}_2 \cdot \text{dm}^{-3}$ ) were also lower, and total hardness was higher ( $118.6 \text{ mg CaCO}_3 \cdot \text{dm}^{-3}$ ).

Simić (2007) gives similar values of the physicochemical water parameters of the Rešava River and the Božička River in Serbia, where there were *L. fluviatilis* populations. Say et al. (1983) noted the results of various

authors concerning the pH and electrolytic conductivity of water from the habitats of *F. antipyretica* and *Ranunculus* sp. (*R. aquatilis* and *R. penicilatus*). The pH from these sites ranged from 6.5 to 8.1, and the electrolytic conductivity was higher than in the Biała Łądecka River –  $70\text{--}615 \mu\text{S} \cdot \text{cm}^{-1}$ . During their research Say et al. (1983) determined the parameters of water in the places where *F. antipyretica* was found. At these places, the pH was 7.5 and electrolytic conductivity was  $296 \mu\text{S} \cdot \text{cm}^{-1}$ .

Metal concentrations in the water of the Biała Łądecka River differed from the concentrations described by Kozubek et al. (2002). They reported higher concentrations of Pb ( $30.69 \mu\text{g} \cdot \text{dm}^{-3}$ ), Hg ( $3.1 \mu\text{g} \cdot \text{dm}^{-3}$ ), Ni ( $5.11 \mu\text{g} \cdot \text{dm}^{-3}$ ), and Zn ( $28.65 \mu\text{g} \cdot \text{dm}^{-3}$ ). The values they reported are several times higher than those in the Biała Łądecka River, however, the Cu content was comparable ( $2.69 \mu\text{g} \cdot \text{dm}^{-3}$ ), though the Cd content was less comparable ( $0.3 \mu\text{g} \cdot \text{dm}^{-3}$ ). In the mountain river Bystrzyca and its tributaries (also in the Sudety Mts), Senze et al. (2007) found higher amounts of Cu, Ni and Zn (for Cu and Zn – 2 times higher), and lower Cd and Pb than we had found in the Biała Łądecka River. On habitats of *F. antipyretica*, Say et al. (1983) found similar amounts of Pb ( $4.2 \mu\text{g} \cdot \text{dm}^{-3}$ ), smaller amounts of Cd ( $0.4 \mu\text{g} \cdot \text{dm}^{-3}$ ), higher amounts of Cu ( $4.5 \mu\text{g} \cdot \text{dm}^{-3}$ ) and significantly higher amounts of Zn ( $83 \mu\text{g} \cdot \text{dm}^{-3}$ ). Samecka-Cymerman et al. (1999) analyzed bryophytes from mountain rivers in basaltic areas of central France. They found different levels of metals in the water: much higher of Cr ( $5.1\text{--}12.9 \mu\text{g} \cdot \text{dm}^{-3}$ ), Ni ( $6.4\text{--}29 \mu\text{g} \cdot \text{dm}^{-3}$ ), and Zn ( $27\text{--}48 \mu\text{g} \cdot \text{dm}^{-3}$ ), Cu which was comparable ( $1\text{--}4.6 \mu\text{g} \cdot \text{dm}^{-3}$ ), and significantly lower of Cd ( $0.003\text{--}0.142 \mu\text{g} \cdot \text{dm}^{-3}$ ) than the same metal levels in the Biała Łądecka River.

The metal concentrations found in the water of the Biała Łądecka River are similar to the concentrations found in other clean European rivers (Neal et al. 1998, Vázquez et al. 2000, Samecka-Cymerman et al. 2005, Gecheva et al. 2011).

The metal content in sediment was significantly higher than that found in the water of the Biała Łądecka River. The metal concentration in sediment to metal concentration in

**Table 5.** Concentrations of Pb, Hg, Cd, Ni, Cr, Cu and Zn in *R. aquatilis*, *F. antipyrretica* and *L. fluviatilis* (mg · kg<sup>-1</sup> in dry weight)

		Pb	Hg	Cd	Ni	Cr	Cu	Zn
<i>Ranunculus aquatilis</i>	Average	77.44 <sup>a</sup>	0.068 <sup>A</sup>	2.560 <sup>A</sup>	10.25 <sup>A</sup>	5.289 <sup>A</sup>	23.46 <sup>A</sup>	701.1 <sup>A</sup>
	Min	8.63	0.027	0.680	5.59	2.453	16.86	148.9
	Max	245.65	0.126	4.460	18.61	9.385	34.11	2093.4
	SD	74.12	0.024	1.093	4.22	2.381	5.59	604.3
<i>Fontinalis antipyrretica</i>	Average	35.66 <sup>ab</sup>	0.104 <sup>B</sup>	0.689 <sup>B</sup>	11.95 <sup>A</sup>	6.446 <sup>A</sup>	23.76 <sup>A</sup>	160.3 <sup>B</sup>
	Min	9.10	0.045	0.243	6.78	2.077	9.34	71.8
	Max	76.95	0.149	1.146	17.68	10.042	41.90	257.4
	SD	20.68	0.035	0.283	3.89	2.319	12.23	74.8
<i>Lemanea fluviatilis</i>	Average	25.07 <sup>b</sup>	0.017 <sup>C</sup>	0.289 <sup>B</sup>	4.98 <sup>B</sup>	1.735 <sup>B</sup>	11.29 <sup>B</sup>	113.6 <sup>B</sup>
	Min	8.99	0.010	0.190	2.91	0.830	6.67	79.7
	Max	51.20	0.025	0.428	8.20	2.732	15.94	145.4
	SD	12.43	0.005	0.085	1.72	0.531	2.64	25.0

different letters indicate statistically significant differences, ab –  $p \leq 0.05$ , AB –  $p \leq 0.01$ ; SD – standard deviation

**Table 6.** Bioconcentration factors (BCF) for Pb, Hg, Cd, Ni, Cr, Cu and Zn in *R. aquatilis*, *F. antipyrretica* and *L. fluviatilis*

		Pb	Hg	Cd	Ni	Cr	Cu	Zn
<i>Ranunculus aquatilis</i>	Average	17851 <sup>a</sup>	149 <sup>A</sup>	4496 <sup>A</sup>	9084 <sup>A</sup>	5523 <sup>A</sup>	13224 <sup>a</sup>	97574 <sup>a</sup>
	Min	1952	39	755	3004	1950	4904	11134
	Max	58197	319	12038	37538	11731	34108	309482
	SD	17483	92	3746	8461	3688	6624	91435
<i>Fontinalis antipyrretica</i>	Average	8047 <sup>a</sup>	226 <sup>A</sup>	920 <sup>B</sup>	26404 <sup>Ba</sup>	12258 <sup>B</sup>	13789 <sup>a</sup>	33978 <sup>b</sup>
	Min	3838	65	280	3601	4096	4092	9195
	Max	21066	410	1432	67776	20766	32404	95321
	SD	5218	123	456	23667	5774	10144	29803
<i>Lemanea fluviatilis</i>	Average	7036 <sup>a</sup>	29 <sup>B</sup>	396 <sup>B</sup>	12046 <sup>ABb</sup>	3029 <sup>A</sup>	5857 <sup>b</sup>	24111 <sup>b</sup>
	Min	2362	10	128	1020	838	2996	11892
	Max	13371	43	681	33234	4914	8951	53007
	SD	4187	13	198	11580	1616	2000	14809

different letters indicate statistically significant differences, ab –  $p \leq 0.05$ , AB –  $p \leq 0.01$ ; SD – standard deviation

water ratio was almost 10,000 for Pb, Ni and Zn. For Cu, this ratio was 6600. For Hg, the ratio was quite different; only 53.

In terms of metal content, the bottom sediment of the Biała Łądecka River (Table 4) belongs to class I classification, only the Pb content was typical of the class II classification of bottom sediment (Bojakowska 2001). These values do not differ from the values reported by other researchers for other rivers (Say et al. 1983, Ciszewski 2001, Ibragimow et al. 2013, Pokorny et al. 2013, Jabłońska-Czapla et al. 2014).

The metal content in the water and sediment of the Biała Łądecka River was characterized by high volatility. The volatility can be expressed as a variability coefficient calculated from the formula: Standard Deviation/Average · 100%. For metals in water, the variability coefficient ranged from 41% (for Cu) to 88% (for Ni). The metal content of the sediment had a lower variability coefficient for Ni, Cu, and Zn (23%, 32%, and 29%, respectively), and a higher variability coefficient for Pb and Hg (73% and 100%, respectively).

The variability coefficient for the metal content in plants was lower than the variability coefficient for the metal content in

water and sediment. However, the variation coefficient differed depending on the plant species. The highest was for *R. aquatilis* (from 24% to 96% for Cu and Pb, respectively) and the lowest was for *L. fluviatilis* (from 22% to 50% for Zn and Pb, respectively).

The metal concentrations in all the species of plants from the Biała Łądecka River study, can be arranged as follows: Hg < Cd < Cr < Ni < Cu < Pb < Zn, and arrangement is the same for the water. Among the studied plants, the lowest metal concentrations were found in *L. fluviatilis*. Of all species of plants, *F. antipyrretica* contained the highest amount of Hg, Ni, Cr, Cu, and *R. aquatilis* contained the highest amount of Pb, Cd and Zn. We found statistically significant differences (Pb –  $p \leq 0.05$ , all other –  $p \leq 0.01$ ) between the content of each metal in *L. fluviatilis*, and in *R. aquatilis*.

The contents of all metals in *F. antipyrretica* were higher than the metals content in *L. fluviatilis*. The differences for Hg, Ni, Cr, and Cu were statistically significant ( $p \leq 0.01$ ).

As indicated by available literature, the metal levels in plants from the present Biała Łądecka River study were similar to the results of other authors. In *R. aquatilis*, Martinez

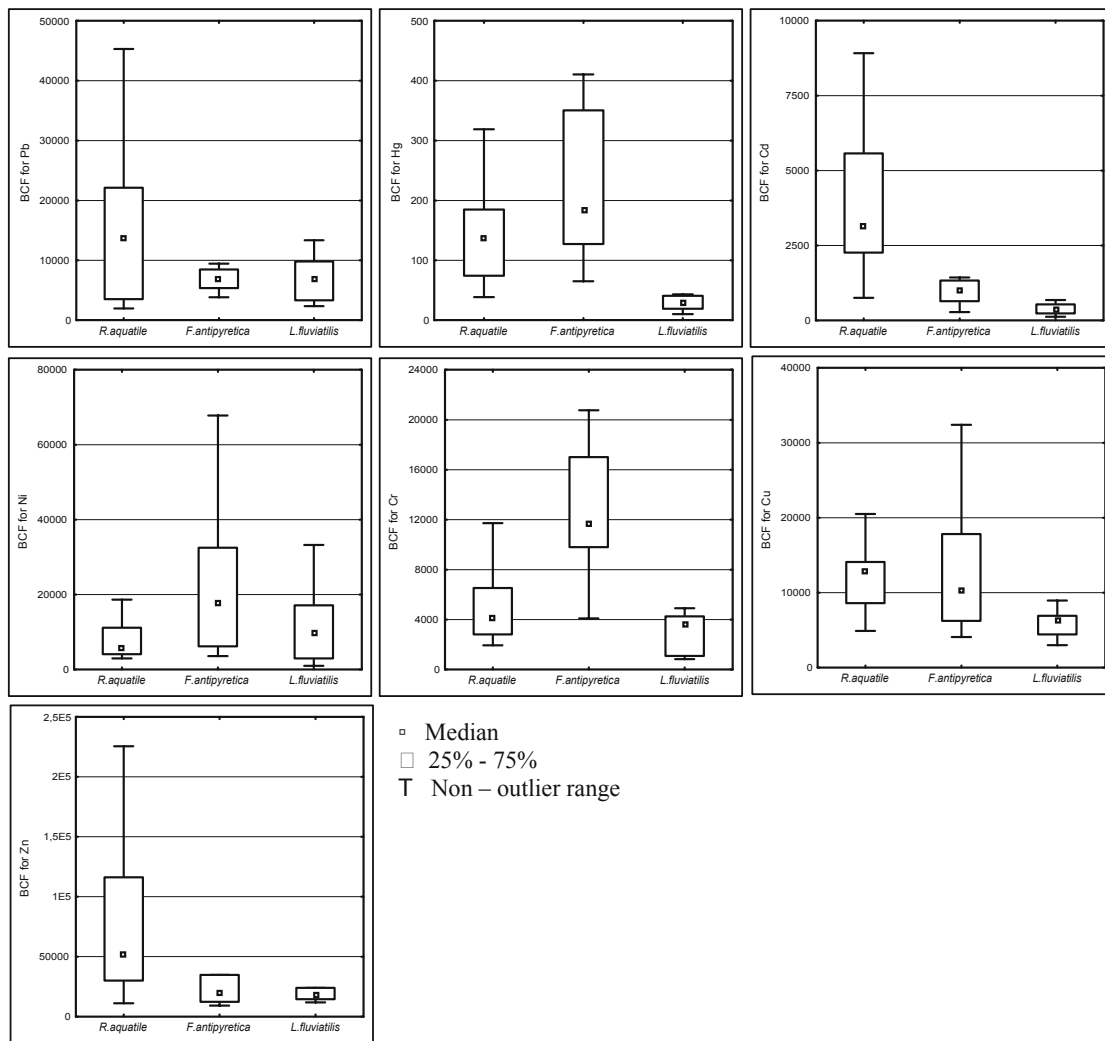


Fig. 1. Median and range of BCF values for metals in *R. aquatilis*, *F. antipyrretica* and *L. fluviatilis*

Table 7. Statistically significant ( $p \leq 0.05$ ) correlations between physicochemical parameters of water and content of the investigated metals in water, sediment and the metal concentration in *R. aquatilis*

Relations between (x and y)	Regression equation	Correlation coefficient (r)	Significant level (p)
Cond : Zn in <i>R. aq.</i>	$y = -268.6133 + 6.7003 x$	0.5084	0.0312
Cond : Cd in <i>R. aq.</i>	$y = -1539.251 + 100.213 x$	0.6955	0.0009
Alk : Ni in <i>R. aq.</i>	$y = 5.4408 + 0.1053 x$	0.5198	0.0226
Th : Zn in <i>R. aq.</i>	$y = -716.4021 + 18.7345 x$	0.5313	0.0233
BOD <sub>5</sub> : Pb in <i>R. aq.</i>	$y = -26.6462 + 36.4196 x$	0.6582	0.0022
Pb in water : Pb in <i>R. aq.</i>	$y = 0.7298 + 14574.4153 x$	0.5512	0.0144
Zn in water : Cd in <i>R. aq.</i>	$y = 1.2915 + 231.4913 x$	0.8333	0.0053
Ni in water : Ni in <i>R. aq.</i>	$y = 6.9403 + 1855.1593 x$	0.6035	0.0062
Cr in water : Zn in <i>R. aq.</i>	$y = 1209.9989 - 5.1465E5 x$	-0.4948	0.0369
Pb in sed. : Pb in <i>R. aq.</i>	$y = -21.8663 + 1.5595 x$	0.8688	0.00000
Zn in sed. : Zn in <i>R. aq.</i>	$y = 1859.2892 - 14.8605 x$	-0.5531	0.0173
Cu in sed. : Zn in <i>R. aq.</i>	$y = 2031.2182 - 92.7053 x$	-0.7159	0.0008
Ni in sed. : Zn in <i>R. aq.</i>	$y = 2251.2317 - 102.9232 x$	-0.5887	0.0102
Hg in sed. : Zn in <i>R. aq.</i>	$y = 1163.0296 - 11629.7886 x$	-0.5988	0.0086

*R. aq.* – *Ranunculus aquatilis*; Cond – electrolytic conductivity; Alk – alkalinity; Th – total hardness; sed. – sediment

**Table 8.** Statistically significant ( $p \leq 0.05$ ) correlations between physicochemical parameters of water and content of the investigated metals in water, sediment and the metal concentration in *F. antipyretica*

Relations between (x and y)	Regression equation	Correlation coefficient (r)	Significant level (p)
Alk : Cr in Font.	$y = 14.0267 - 0.2254 x$	-0.9581	0.00005
Th : Cu in Font.	$y = -45.5121 + 1.0276 x$	0.7118	0.0315
O <sub>2</sub> : Cu in Font.	$y = 219.207 - 16.8977 x$	-0.6875	0.0407
Pb in water : Pb in Font.	$y = 8.4316 + 5117.0033 x$	0.7504	0.0198
Pb in water : Ni in Font.	$y = 6.6759 + 991.1837 x$	0.7730	0.0146
Zn in water : Pb in Font.	$y = 73.4723 - 6065.2731 x$	-0.7263	0.0267
Zn in water : Ni in Font.	$y = 18.5588 - 1060.0333 x$	-0.6751	0.0460
Ni in water : Pb in Font.	$y = 20.4559 + 13036.7567 x$	0.8171	0.0072
Hg in water : Cr in Font.	$y = 3.7307 + 4610.7845 x$	0.6847	0.0419
Pb in sed. : Pb in Font.	$y = 14.8068 + 0.3328 x$	0.8211	0.0067
Pb in sed. : Cu in Font.	$y = 12.0882 + 0.1862 x$	0.7767	0.0138
Pb in sed. : Ni in Font.	$y = 7.6636 + 0.0684 x$	0.8976	0.0010
Zn in sed. : Pb in Font.	$y = -11.9592 + 0.6015 x$	0.7434	0.0217
Hg in sed. : Pb in Font.	$y = 15.5155 + 479.7585 x$	0.7457	0.0211

Font – *Fontinalis antipyretica*; Alk – alkalinity; Th – total hardness; O<sub>2</sub> – dissolved oxygen; sed. – sediment

**Table 9.** Statistically significant ( $p \leq 0.05$ ) correlations between physicochemical parameters of water and content of the investigated metals in water, sediment and the metal concentration in *L. fluviatilis*

Relations between (x and y)	Regression equation	Correlation coefficient (r)	Significant level (p)
Cond : Hg in Lem.	$y = -0.0029 + 0.0002 x$	0.8831	0.0016
Th : Ni in Lem.	$y = -7.4287 + 0.1938 x$	0.6706	0.0481
Th : Cd in Lem.	$y = -0.4263 + 0.0112 x$	0.7812	0.0129
Th : Cr in Lem.	$y = -2.6022 + 0.0677 x$	0.7597	0.0175
Zn in water : Cr in Lem.	$y = 2.8998 + 207.1143 x$	-0.7709	0.0150
Pb in sed. : Pb in Lem.	$y = 15.0003 + 0.1627 x$	0.6733	0.0468
Cu in sed. : Hg in Lem.	$y = 0.0258 - 0.0006 x$	-0.6933	0.0384
Ni in sed. : Hg in Lem.	$y = 0.0304 - 0.0009 x$	-0.7854	0.0121

Lem. – *Lemanea fluviatilis*; sed. – sediment; Cond – electrolytic conductivity; Th – total hardness;

**Table 10.** Statistically significant ( $p \leq 0.05$ ) correlations between physicochemical parameters of water and content of the investigated metals in water, sediment and the BCF for metal in *R. aquatilis*

Relations between (x and y)	Regression equation	Correlation coefficient (r)	Significant level (p)
Zn in water : BCF Pb in R. aq.	$y = 3456.1036 + 1.6003E6 x$	0.62	0.004
Pb in water : BCF Cd in R. aq.	$y = 1137.7484 - 1.186E6 x$	-0.75	0.020
Zn in water : BCF Cd in R. aq.	$y = 936.8764 + 6.4978E5 x$	0.68	0.043
Cd in R. aq. : BCF Ni in R. aq.	$y = 18054.0579 - 3598.0412 x$	-0.67	0.047
Cond : BCF Cu in R. aq.	$y = 1539.251 + 100.213 x$	0.70	0.001
Alk : BCF Cu in R. aq.	$y = 2733.5963 + 229.571 x$	0.72	0.001
Th : BCF Cu in R. aq.	$y = -7840.0059 + 274.0981 x$	0.72	0.001
Ni in water : BCF Cu in R. aq.	$y = 9001.8029 + 2.3663E6 x$	0.49	0.033
Cr in water : BCF Cu in R. aq.	$y = 18950.3289 - 5.7569E6 x$	-0.49	0.033
Hg in sed. : BCF Cu in R. aq.	$y = 17272.3509 - 1.0657E5 x$	-0.50	0.029

R. aq. – *Ranunculus aquatilis*; Cond – electrolytic conductivity; Alk – alkalinity; Th – total hardness; sed. – sediment

**Table 11.** Statistically significant ( $p \leq 0.05$ ) correlations between physicochemical parameters of water and content of the investigated metals in water, sediment, and plant and the BCF for metal in *F. antipyretica*

Relations between (x and y)	Regression equation	Correlation coefficient (r)	Significant level (p)
Hg in water : BCF Pb in Font.	$y = 814.1135 + 1.2283E7 x$	0.81	0.008
Th : BCF Hg in Font.	$y = -468.5711 + 10.2205 x$	0.68	0.030
O <sub>2</sub> : BCF Hg in Font.	$y = 2602.9746 - 205.9489 x$	-0.80	0.006
Hg in sed. : BCF Ni in Font.	$y = 47120.207 - 4.9324E5 x$	-0.67	0.048
Th : BCF Cu in Font.	$y = -52677.9782 + 985.9972 x$	0.82	0.006
O <sub>2</sub> : BCF Cu in Font.	$y = 2.1071E5 - 17024.9992 x$	-0.84	0.005
Pb in sed. : BCF Zn in Font.	$y = 3880.3069 + 480.2388 x$	0.82	0.007
Zn in sed. : BCF Zn in Font.	$y = -31942.6097 + 832.6085 x$	0.71	0.031

Font – *Fontinalis antipyretica*; Th – total hardness; O<sub>2</sub> – dissolved oxygen; sed. – sediment

**Table 12.** Statistically significant ( $p \leq 0.05$ ) correlations between physicochemical parameters of water and content of the investigated metals in water, sediment, and plant and the BCF for metal in *L. fluviatilis*

Relations between (x and y)	Regression equation	Correlation coefficient (r)	Significant level (p)
Cd in water : BCF Pb in Lem.	$y = 13153.7168 - 6.7147E6 x$	-0.80	0.009
Alk : BCF Cd in Lem.	$y = 1010.384 - 18.8766 x$	-0.71	0.033
Cu in water : BCF Cd in Lem.	$y = 697.097 - 1.4022E5 x$	-0.68	0.045
Hg in sed. : BCF Ni in Lem.	$y = 22830.9852 - 2.5959E5 x$	-0.73	0.025
BOD <sub>5</sub> : BCF Cu in Lem.	$y = 12016.3846 - 2399.7702 x$	-0.77	0.015
Cd in water : BCF Cu in Lem.	$y = 8658.223 - 3.0745E6 x$	-0.77	0.015
Pb in sed. : BCF Zn in Lem.	$y = 10519.1522 + 219.5411 x$	0.76	0.017
Zn in sed. : BCF Zn in Lem.	$y = -6254.2987 + 388.7518 x$	0.71	0.034

Lem. – *Lemanea fluvitans*; Alk – alkalinity; sed. – sediment

et al. (2011) found higher amounts of Hg and Cu (respectively, 1.045 mg · kg<sup>-1</sup> and 53.15 mg · kg<sup>-1</sup>) but smaller amounts of Cd and Zn (respectively, 0.555 mg · kg<sup>-1</sup> and 61.25 mg · kg<sup>-1</sup>).

Samecka-Cymerman et al. (1996) found in *R. aquatilis* (from contaminated river) higher amounts of Hg and Ni (0.16 mg · kg<sup>-1</sup> and 9.5 mg · kg<sup>-1</sup>) but smaller Pb, Cd, Cr, Cu and Zn (1.6 mg · kg<sup>-1</sup>, 0.63 mg · kg<sup>-1</sup>, 2.4 mg · kg<sup>-1</sup>, 7.1 mg · kg<sup>-1</sup> and 101 mg · kg<sup>-1</sup>).

There is much more information in the literature concerning *F. antipyretica*. In our study, the indicated concentrations of Pb, Cd, Ni, Cr, and Cu in this moss were in the range reported by the following authors (Say et al. 1983, Vazquez et al. 2004, Gapeeva et al. 2010, Gecheva et al. 2011, Martinez et al. 2011). However, in the case of Hg and Zn, the maximum values given by the authors are much higher than those obtained in our study. In *F. antipyretica*, Vazquez et al. (2004) even found 0.76 mg · kg<sup>-1</sup> (5 times more than the maximum value found in our study) and 1107 mg · kg<sup>-1</sup> Zn (4 times more). In *F. antipyretica*, Say et al. (1983) found 2825 mg · kg<sup>-1</sup> Zn (11 times more).

Correlation coefficients between the analyzed factors (physical and chemical parameters of water, metal concentration in water, sediment, and plants) were calculated. Among the statistically proven correlations, some can be distinguished. What would seem to be the obvious relationship of the metal content in water and the content of the same metal in plant,

was actually only statistically confirmed in three cases. The three cases were: 1) Pb in water and Pb in *R. aquatilis*, and 2) Pb in water vs. Pb and *F. antipyretica*, and 3) Ni in water and Ni in *R. aquatilis*. All these correlations are positive and strong. Statistically significant correlations between a specific metal content in sediment and that specific metal content in plant were equally rare. The correlations were statistically confirmed only for Pb in sediment and Pb in all of plant species, and the correlations were statistically confirmed for Zn in sediment and Zn in *R. aquatilis*. Interestingly, with an increase of Zn content in sediment, there was a decrease in the concentration of Zn in *R. aquatilis* ( $r = -0.55$ ). Samecka-Cymerman et al. (1999) have confirmed other correlations: Cu in water and Ni in bryophytes, Cr in water and Pb in bryophytes, Pb in water and Cd in bryophytes. But their research was carried out in different water conditions, where, above all, there was a low pH (5.7–6.3).

Say et al. (1983) demonstrated a correlation between Cu and Zn in water and in *F. antipyretica*. They also showed a correlation between the contents of Cd, Pb, Cu, and Zn in sediment and in *F. antipyretica*.

The occurrence of a positive correlation between the content of the metal in water and its concentration in plant may indicate that the metal is absorbed by green part of plant, which is washed by water. The lack of such correlation does not necessarily imply a different way of metal absorption. It



may be due to the fact that metal is present in water in less bioavailable forms. Positive correlations between the metal content in sediment and metal concentrations in plant could indicate metal absorption by the roots or rhizoids.

Bioconcentration factors vary significantly depending on the metal and the plant species. The calculated BCFs for metals, arranged in ascending order, are as follows:

<i>Ranunculus aquatilis</i>	Hg < Cd < Cr < Ni < Cu < Pb < Zn
<i>Fontinalis antipyretica</i>	Hg < Cd < Pb < Cr < Cu < Ni < Zn
<i>Lemanea fluviatilis</i>	Hg < Cd < Cr < Cu < Pb < Ni < Zn

Among the analyzed metals, plants accumulated the highest amount of Zn, and the least of Cd and Hg. For nickel, the lowest BCF was calculated in *R. aquatilis*, for all other metals the lowest BCF was calculated in *L. fluviatilis*.

In conclusion, we can say that the lowest bioaccumulation of metals was characterized by *L. fluviatilis*, and the highest by *F. antipyretica*.

Physicochemical parameters may have statistically significant influence ( $p \leq 0.05$ ) on the BCFs for metals in plants. The following statistically significant relationships were confirmed: BCF for Cu in *R. aquatilis* was correlated with the alkalinity, electrolitical conductivity, and total hardness; BCF for Cu in *F. antipyretica* was correlated with total hardness and dissolved oxygen; and BCF for Cu in *L. fluviatilis* was correlated with BOD<sub>5</sub>. The BCF for Hg in *F. antipyretica* was correlated with total hardness and dissolved oxygen. The BCF for Cd in *L. fluviatilis* was correlated with alkalinity.

We managed to see some significant trends. The bioconcentration factor for Zn in *F. antipyretica* and *L. fluviatilis* was positively ( $r > 0.7$ ) correlated with the contents of Pb and Zn in river sediments. The bioconcentration factor for Ni was inversely proportional to concentrations of Hg in the sediment, for *F. antipyretica*, and *L. fluviatilis*. Whereas, in the case of *R. aquatilis*, Hg in sediment correlated (also negatively) with BCF for Cu.

## Conclusions

1. The lowest concentrations of Hg and the highest concentrations of Zn were determined in the environmental elements of the Biała Łądecka River.
2. In ascending order, the metal concentrations were: water < sediments < plants.
3. The concentrations of metals in the water and in the plants can be arranged as follows: Hg < Cd < Cr < Ni < Cu < Pb < Zn.
4. Out of three plants species, the lowest concentrations of all metals were determined in *L. fluviatilis*.
5. In all plants species, the lowest BCFs were calculated for Hg, and the highest for Zn.

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## References

- Atherton, I., Bosanquet, S. & Lawley, M. (2010). *Mosses and Liverworts of Britain and Ireland: a field guide*, British Bryological Society. Latimer Trend & Co. Ltd., Plymouth, UK 2010.
- Bojakowska, I. (2001). Criteria for assessing pollution of water sediment, *Przegląd Geologiczny*, 49, pp. 213–218. (in Polish)
- Bory de Saint-Vincent, J.B. (1808). Mémoire sur le genre Lemanea de la famille des Conferves, *A paper on the genus Lemanea of the family Conferva*, 12, pp. 177–190. (in French)
- Ciszewski, D. (2001). Flood-related changes in heavy metal concentrations within sediments of the Biała Przemsza River, *Geomorphology*, 40, pp. 205–218.
- Cook, C.D.K. (1963). Studies in Ranunculus subgenus Batrachium (D.C) A. Gray. II. General morphological consideration in the taxonomy of the subgenus, *Watsonia*, 5, pp. 294–303.
- Dahlgren, G. (1991). Karyological investigations in Ranunculus subg. Batrachium (Ranunculaceae) on the Aegean Islands, *Plant Systematics and Evolution*, 177, pp. 193–211.
- Gapeeva, M. V., Dolotov, A. V. & Chemeris, E. V. (2010). Prospects of using mosses (*Fontinalis antipyretica* Hedw. and *Pylaisia polyantha* (Hedw.) Bruch et al.) as indicators of environmental contamination with heavy metals, *Russian Journal of Ecology*, 41, pp. 28–31.
- Gecheva, G., Yurukova, L. & Ganeva, A. (2011). Assessment of pollution with aquatic bryophytes in Maritsa River (Bulgaria), *Bulletin of Environmental Contamination and Toxicology*, 87, pp. 480–485.
- Ibragimow, A., Walna, B. & Siepak, M. (2013). Physico-chemical parameters determining the variability of actually and potentially available fractions of heavy metals in fluvial sediments of the Middle Odra River, *Archives of Environmental Protection*, 39, pp. 3–16.
- Jabłońska-Czapla, M., Szopa, S. & Rosik-Dulewska, C. (2014). Impact of mining dump on the accumulation and mobility of metals in the Bytomka River Sediments, *Archives of Environmental Protection*, 40, pp. 3–19.
- Kabata-Pendias, A. & Pendias, H. (1993). *Biogeochemistry of trace elements*. Wydawnictwo Naukowe PWN, Warszawa 1993. (in Polish)
- Kabata-Pendias, A. & Pendias, H. (1993). *Biogeochemistry of trace elements*. Wydawnictwo Naukowe PWN, Warszawa 1993. (in Polish)
- Kozubek, M. & Marek, J. (2002). Heavy metals in Bystrzyca Dusznicka River and its tributaries, *Zeszyty Naukowe AR we Wrocławiu*, 447, pp. 89–99. (in Polish)
- Kučera, P. & Marvan, P. (2004). Taxonomy and distribution of Lemanea and Paralemanea (Lemaneaceae, Rhodophyta) in the Czech Republic, *Preslia, Praha*, 76, pp. 163–174.
- Martinez, E.A. & Shu-Nyamboli, C. (2011). Determination of selected heavy metal concentrations and distribution in a southwestern stream using macrophytes, *Ecotoxicology and Environmental Safety*, 74, pp. 1504–1511.
- Neal, C., Robson, A.J., Wass, P., Wade, A.J., Ryland, G.P., Leach, D.V. & Leeks, G.J.L. (1998). Major, minor, trace element and suspended sediment variations in the River Derwent, *The Science of the Total Environment*, 210–211, pp. 163–172.
- Pokorný, P., Senze, M., Dobicki, W., Kowalska-Górska, M. & Polechoński, R. (2013). Geochemical assessment of Western Pomerania watercourses, *Przemysł Chemiczny*, 92, pp. 1768–1771. (in Polish)
- Samecka-Cymerman, A. & Kempers, A.J. (1996). Bioaccumulation of heavy metals by aquatic macrophytes around Wrocław, Poland, *Ecotoxicology and Environmental Safety*, 35, pp. 242–247.
- Samecka-Cymerman, A. & Kempers, A.J. (1999). Background concentrations of heavy metals in aquatic bryophytes used

- for biomonitoring in basaltic areas (a case study from central France), *Environmental Geology*, 39, pp. 117–122.
- Samecka-Cymerman, A., Kolon, K. & Kempers, A.J. (2005). A comparison of native and transplanted *Fontinalis antipyretica* Hedw. as biomonitors of water polluted with heavy metals, *Science of the Total Environment*, 341, pp. 97–107.
- Say, P.J. & Whitton, B.A. (1983). Accumulation of heavy metals by aquatic mosses. 1: *Fontinalis antipyretica* Hedw., *Hydrobiologia*, 100, pp. 245–260.
- Senze, M., Kowalska-Górska, M. & Pokorny, P. (2009). Metals in chosen aquatic plants in a lowland dam reservoir, *Journal of Elementology*, 14, pp. 147–156.
- Senze, M., Pokorny, P., Polechoński, R., Dobicki, W., Kowalska-Górska, M. & Bednarska, M. (2007). Heavy metals in hydromacrophytes of the Bystrzyca River and its tributaries above the dammed reservoir of Lubachów, *Chemistry for Agriculture*, 8, pp. 227–233.
- Siemińska, J. (2006). Red list of the algae in Poland, in: *Red list of plants and fungi in Poland*, Mirek, Z., Zarzycki, Z., Wojewoda, W. & Szelański, Z., (Eds.). Szafer Institute of Botany, Polish Academy of Sciences, Kraków 2006.
- Simić, S. (2007). Morphological and ecological characteristics of rare and endangered species *Lemanea fluviatilis* (Linné) C. Ag. (Lemnaceae, Rhodophyta) on new localities in Serbia, *Kragujevac Journal of Science*, 29, pp. 97–106.
- Starmach, K. (1977). Phaeophyta – Brunatnice, Rhodophyta – Krasnorosty, in: *Flora słodkowodna Polski*, Starmach, K. & Siemińska, J. (Eds.). Polska Akademia Nauk, Warszawa & Kraków 1977. (in Polish)
- Vázquez, M.D., Fernández, J.A., López, J. & Carballeira, A. (2000). Effects of water acidity and metal concentration on accumulation and within-plant distribution of metals in the aquatic bryophyte *Fontinalis antipyretica*, *Water, Air, and Soil Pollution*, 120, pp. 1–19.
- Vázquez, M.D., Wappelhorst, O. & Markert, B. (2004). Determination of 28 elements in aquatic moss *Fontinalis antipyretica* Hedw. and water from the upper reaches of the river Nysa (Cz, D) by ICP – MS, ICP – OES and AAS, *Water, Air, and Soil Pollution*, 152, pp. 153–172.
- Vis, M.L. & Sheath, R.G. (1992). Systematics of the freshwater red algal family Lemnaceae in north America, *Phycologia*, 31, pp. 164–179.

## Bioakumulacja metali ciężkich w makrofitach submersyjnych górskiej rzeki Białej Łądeckiej (Polska, Sudety)

**Streszczenie:** Badania przeprowadzono na górskiej rzece Białej Łądeckiej. Pod względem hydromorfologicznym i zagospodarowania zlewni jest ona porównywalna z wieloma europejskimi rzekami górkimi. Celem pracy było określenie współczynników biokoncentracji metali ciężkich (Pb, Cd, Hg, Ni, Cr, Cu i Zn) w *Ranunculus aquatilis* (L.) Dumort., *Fontinalis antipyretica* (L. ex Hedw.) i *Lemanea fluviatilis* (L.) C.Ag. Określono zawartość metali w wodzie, osadzie, i roślinach zanurzonych. Stężenia metali w roślinach można uszeregować w następujący sposób: Hg < Cd < Cr < Ni < Cu < Pb < Zn. Najwyższe stężenia Hg, Ni, Cr i Cu stwierdzono w *F. antipyretica*, a najwyższe stężenie Pb, Cd i Zn w *R. aquatilis*. Stężenia metali w *L. fluviatilis* zawsze były najniższe. Na podstawie zawartości metalu w wodzie i w roślinach obliczono współczynniki biokoncentracji (BCFs). Współczynniki BCFs były najniższe dla *L. fluviatilis* a najwyższe dla *F. antipyretica*. Spośród analizowanych metali, rośliny kumulowały najwięcej Zn a najmniej Hg. BCFs dla Zn wynosiło od 24111 (w *L. fluviatilis*) do 97574 (w *R. aquatilis*), a BCFs dla Hg wynosiło od 29 (w *L. fluviatilis*) do 226 w (*F. antipyretica*).